



RESEARCH ARTICLE

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Kev Points:

- Observations show robust correlations and consistent linear regression slopes of O₃-CH₂O, O₃-CO and CO-CH₂O
- Biogenic isoprene oxidation makes the largest contribution to the regression slope of O₃-CH₂O across much of the eastern United States
- We build a fast-response O₂ estimator using input near-surface CH₂O and CO concentrations to improve surface O₃ distribution monitoring

Supporting Information:

- Supporting Information S1
- Figure S1

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United States in Summer

Monoxide Concentrations Over the Eastern

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Estimator of Surface Ozone Using Formaldehyde and Carbon

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Abstract Strong correlations of O_3 -CH₂O, O_3 -CO and CO-CH₂O were observed during the Deriving Information on Surface Conditions from Column and Vertically Resolved Observations Relevant to Air Quality (DISCOVER-AQ) aircraft experiment in July 2011 over the Washington-Baltimore area. The linear regression slopes of observed O₃-CH₂O, O₃-CO, and CO-CH₂O do not vary significantly with time (11 a.m. to 4 p.m.) or altitude in the boundary layer. These observed relationships are simulated well by a regional chemical transport model. Using tagged-tracer simulations, we find that biogenic isoprene oxidation makes the largest contribution to the regression slope of O_3 -CH₂O across much of the eastern United States, providing a good indicator for O_3 enhanced by biogenic isoprene oxidation. In contrast, the regression slope of O_3 -CO is controlled by both anthropogenic and biogenic emissions. Therefore, we use the CO-CH₂O relationship to separate biogenic from anthropogenic contributions to CO. By combining these regressions, we can track the contributions to surface O₃ by anthropogenic and biogenic factors and build a fast-response ozone estimator using near-surface CH_2O and CO concentrations as inputs. We examine the quality of O_3 estimator by increasing or decreasing anthropogenic emissions by up to 50%. The estimated O₃ distribution is in reasonably good agreement with the full-model simulations ($R^2 > 0.77$ in the range of -30% to +50% of anthropogenic emissions). The analysis provides the basis for using high-quality geostationary satellites with UV, thermal infrared, or near-infrared instruments for observing CH₂O and CO to improve surface O₃ distribution monitoring. The estimation model can also be applied to derive observation-derived regional metrics to evaluate and improve full-fledged 3-D air quality models.

1. Introduction

Ozone (O₃) is a major pollutant in the troposphere (e.g., Lelieveld & Dentener, 2000; Logan et al., 1981; Wang et al., 1998; Wang & Jacob, 1998). Thus, monitoring tropospheric O₃ at regional and global scales is important for environmental protection. Spaceborne remote sensing utilizing its absorption features in the ultraviolet (UV) and the thermal infrared (TIR) bands is the most convenient way to provide O_3 spatial distributions around the globe. However, because of the molecular scattering of UV (X. Liu, Bhartia, et al., 2010) and lack of contrast of TIR (Beer, 2006), the satellite observations for O_3 still show a limited sensitivity in the lowermost troposphere, especially near the surface, which is directly relevant to air quality (e.g., Cuesta et al., 2013).

Three-dimensional air quality models can provide information on the distribution of surface O₃, but the accuracy of surface O₃ simulations is limited by uncertainties in precursor emissions, atmospheric processes, and nonlinear photochemistry. Some model uncertainties can be mitigated through probabilistic approaches (e.g., Dabberdt et al., 2004; Delle Monache, Deng, et al., 2006; Delle Monache, Hacker, et al., 2006; Vautard et al., 2009). Statistical methods have also been applied to estimate O_3 distributions, including classification and regression trees, linear regression, and neural networks (e.g., Biancofiore et al., 2015; Burrows et al., 1995; Cobourn, 2007; Perez & Reyes, 2006; Shad et al., 2009; Van der Wal & Janssen, 2000). The advantage of such statistical modeling is that it offers moderate to high accuracy at a moderate cost (Y. Zhang et al., 2012). However, the nature of statistical modeling often requires a suite of input variables and does not enable better understanding of chemical and physical processes (e.g., Guillas et al., 2008).

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An alternative to numerical or statistical modeling is to make use of the observations of O_3 surrogates. The correlations between O_3 and other chemical species might be used effectively to diagnose O_3 chemical and physical processes (e.g., Chin et al., 1994; Koo et al., 2012; Parrish et al., 1993; Wang & Zeng, 2004). Carbon monoxide (CO) is often observed and simulated to have a linear relationship with O3 in the lower atmosphere (e.g., Buhr et al., 1996; Cardenas et al., 1998; Cheng et al., 2017; Chin et al., 1994; Cooper, Moody, Parrish, Trainer, Holloway, et al., 2002; Cooper, Moody, Parrish, Trainer, Ryerson, et al., 2002; Fishman & Seiler, 1983; Honrath et al., 2004; Huntrieser et al., 2005; Q. Li et al., 2002; Mao & Talbot, 2004; Parrish et al., 1993, 1998). Over the eastern United States, the observed slope of O₃ to CO at ~0.3 reflects contributions by CO from primary anthropogenic emissions and that from biogenic isoprene oxidation (Cheng et al., 2017) due in part to the production of both O_3 and CO from biogenic volatile organic compound (VOC) oxidation (Atkinson & Arey, 1998; Choi et al., 2010; Geng et al., 2011; Guenther et al., 1995; Hudman et al., 2008; K.-Y. Lee et al., 2014; Pang et al., 2009; Pierce et al., 1998; Y. Zhang & Wang, 2016). It implies that the observations of CO concentrations can be potentially applied to track the contributions to surface O_3 by anthropogenic and biogenic factors. In addition to Environmental Protection Agency surface monitoring networks, satellite observations of lower tropospheric CO are more promising than O₃ over polluted regions since CO concentrations are usually higher in the boundary layer than the free troposphere and unlike O₃ it does not have high concentrations in the stratosphere, although both CO and O₃ have substantial free tropospheric columns relative to the boundary layer. CO can be detected by satellite TIR, near-infrared (NIR), and joint TIR and NIR instruments, such as the NIR SCanning Imaging Absorption spectroMeter for Atmospheric CHartographY (SCIAMACHY; e.g., De Laat et al., 2012), the Measurement of Pollution in the Troposphere (e.g., Emmons et al., 2004; Straume et al., 2005), and Atmospheric Infrared Sounder Aumann et al., 2003).

Formaldehyde (CH₂O) is a principal intermediate species in the oxidation of atmospheric hydrocarbons (e.g., Duane et al., 2002; Fried et al., 2011; Pang et al., 2009; Wiedinmyer et al., 2001, and references therein). It is also a major radical source leading to ozone production in the presence of nitrogen oxides (NO_x; e.g., Z. Liu, Wang, Vrekoussis, et al., 2012). We will show that regional CH₂O is also correlated to O₃ in section 3. Since the sources of CO and CH₂O are often different, the two correlations can provide separate constraints on O₃ distributions. While regulatory monitoring of surface CH₂O is unavailable, CH₂O is detectable from space with good sensitivities in the boundary layer by measuring backscattered solar UV between 325 and 360 nm (Chance et al., 2000), including SCIAMACHY (Wittrock et al., 2006), Ozone Monitoring Instrument (González Abad et al., 2015; Kurosu et al., 2004), Global Ozone Monitoring Experiment-2 (GOME-2; De Smedt et al., 2012), and Ozone Mapping Profiler Suite (González Abad et al., 2016; C. Li et al., 2015). The total uncertainty of the HCHO vertical column data is typically in the range of 50–105% for each measurement. Through averaging, the uncertainties for monthly means are down to 20–40% for GOME-2A and SCIAMACHY (De Smedt et al., 2008), 38% for Ozone Monitoring Instrument, and 46% for GOME-2B (Zhu et al., 2016). Therefore, the uncertainty of the monthly HCHO vertical column data from polar-orbiting instruments are in the range of 20–50%.

In this study, we apply a 3-D chemical transport model to quantitatively study factors contributing to the observed correlations and regression slopes of O₃ with CH₂O and CO and those of CO with CH₂O using the Deriving Information on Surface Conditions from Column and Vertically Resolved Observations Relevant to Air Quality (DISCOVER-AQ) measurements. The analysis will show that it is feasible to use observed aircraft CO and CH₂O concentrations to improve estimates of the surface O₃ distribution over the eastern United States. The uncertainties of current satellite-derived near-surface CO and CH₂O are still too large (e.g., Buchwitz et al., 2007; Gloudemans et al., 2005; L. Zhang, Jiang, et al., 2016) to be applied in this method. The Environmental Protection Agency surface CO monitoring data are also unusable since CO concentrations are often below the reporting limit (e.g., Zeng & Wang, 2011). We therefore use the model simulated data, which reasonably captures the observed aircraft concentrations and relationships of O_3 , CO, and CH₂O during the DISCOVER-AQ campaign, to develop and evaluate a surface O₃ estimation model with surface CO and CH₂O as input parameters. In the future, geostationary satellite instruments such as Tropospheric Emissions: Monitoring of Pollution (TEMPO; Chance et al., 2013) and Geostationary Carbon Observatory (Polonsky et al., 2014) will greatly improve the monitoring of near-surface measurements of O₃ precursors with sufficient accuracy and therefore make it possible to use observed CH₂O and CO in the O3 estimator we develop here. We describe the 3-D chemical transport model, DISCOVER-AQ data set,



analysis methods, the O_3 estimation model, and validation method in section 2. Section 3 describes the analysis, modeling, and evaluation results. Discussion of implementing the O_3 estimator and conclusions are given in sections 4 and 5, respectively.

2. Data and Analysis Methods

2.1. Observations and Simulations

The observation data used in this study were obtained from the NASA 2011 DISCOVER-AQ airborne campaign (http://www-air.larc.nasa.gov/missions/discover-aq/discover-aq.html). Sampling by the NASA P-3B aircraft was conducted from Washington's Beltway northeast to Baltimore and continuing on to the Delaware state line and occasionally over the Chesapeake Bay. Fourteen flights over six locations selected for aircraft spirals were carried out to measure the vertical structures of pollutants. Two hundred fifty-three daytime vertical profiles from 300 m to 5 km were measured between 27 June and 31 July. CO was measured by a diode laser spectrometer (Sachse et al., 1987). O₃ was measured by the National Center for Atmospheric Research four-channel chemiluminescence instrument (Weibring et al., 2010). The uncertainties of the measurements on these two species are 2% and 5%, respectively. CH_2O was measured by a difference frequency generation absorption spectrometer (Weibring et al., 2010). For CH_2O levels above 1 ppbv the total measurement uncertainty at the 1 σ level was estimate to be around 5%, which folds in systematic and limits of detection uncertainties. To evaluate model simulations with the observations, we identify the model profiles corresponding to the locations of aircraft spirals and the time of aircraft sampling. Corresponding model vertical profiles and observations are used in correlation analysis to evaluate model performance.

We use a 3-D Regional chEmical trAnsport Model (REAM) to represent the observations and then conduct further correlation analyses. The REAM model was applied in previous studies to analyze vertical mixing, large-scale transport, emission estimates, and tropospheric chemistry over North America and East Asia (e.g., Cheng et al., 2017; Choi et al., 2005; Choi, Wang, Yang, et al., 2008; Choi, Wang, Zeng, et al., 2008; Gu et al., 2013, 2014, 2016; Jing et al., 2006; Z. Liu, Wang, et al., 2010; Z. Liu, Wang, Gu, et al., 2012; Z. Liu, Wang, Vrekoussis, et al., 2012; Z. Liu et al., 2014; Wang et al., 2006, 2007; Yang et al., 2011; Zeng et al., 2003, 2006; Zhao & Wang, 2009; Zhao, Wang, Choi, & Zeng, 2009; Zhao, Wang, & Zeng, 2009; Zhao et al., 2010; Y. Zhang, Wang, et al., 2016; Y. Zhang & Wang, 2016; R. Zhang et al., 2017). The model domain covers the contiguous United States with a horizontal resolution of 36×36 km². The chemistry mechanism in REAM is the GEOS-Chem standard chemical mechanism (V9-02; Bey et al., 2001) with updates of kinetics data (http://jpldataeval.jpl.nasa.gov). The anthropogenic emission inventory used in the model is the 2011 National Emission Inventory (https://www.epa.gov/air-emissions-inventories/2011-national-emissions-inventory-nei-data). The biogenic isoprene emissions are the results of the Model of Emissions of Gases and Aerosols from Nature version 2.1 (Guenther et al., 2012). Initial and boundary conditions for chemical tracers are taken from the GEOS-Chem (V9-02) 2°× 2.5°simulation results (Bey et al., 2001). Meteorology fields are from the Weather Research and Forecasting model, which assimilated products from the Climate Forecast System Reanalysis (http://cfs.ncep.noaa.gov/cfsr).

The previous study by Cheng et al. (2017) shows that REAM simulates well the observed vertical and temporal variations of O₃, CO, NO_x, isoprene, and CH₂O, as well as the correlation between O₃ and CO, and the O₃-CO regression slopes during the 2011 DISCOVER-AQ campaign. In this work, we therefore focus on analyzing the correlation and regression slope of O₃-CH₂O. We trace separately via tagged tracers three different CH₂O sources, primary anthropogenic emissions and the oxidation of anthropogenic VOCs (CH₂O_{anthroVOCs}), the oxidation of biogenic isoprene (CH₂O_{biolSOP}), and transport from model lateral and upper boundaries (CH₂O_{BC}), to analyze the contribution from each source to the observed O₃-CH₂O relationship. Other biogenic VOCs are not taken into account because isoprene provides the source for the vast majority biogenic CH₂O (e.g., Guenther et al., 2012; Kesselmeier & Staudt, 1999; Lathiere et al., 2006; Sindelarova et al., 2014). In tagged-tracer simulations, relevant species and radicals, such as O₃, NO_x, and HO_x (OH and HO₂), are fixed using results archived from the standard simulation. The sum of the three individual tagged tracers is within 2% of the total CH₂O concentrations in the standard simulation for grid cells over the Washington-Baltimore region. We carry out minor scaling adjustments in postprocessing, assuming that relative CH₂O attributions stay the same, to ensure that the sum of the CH₂O tracers is the same as the total CH₂O for each grid cell in the standard simulation. We evaluate the scaling adjustments by calculating the relative error of the sum of



the three tagged CH₂O to the total simulated CH₂O (Figure S1 in the supporting information). In the study domain (mainly the southeast United States), the error is <0.5%. Since the magnitude of total CH₂O is lower in other areas, the error is lager but is still <1.5% for western United States and 3% for other rural regions. With simulated CH₂O attribution results, we can decompose the O₃-CH₂O regression slope into three subslopes of the corresponding CH₂O tracers (equation (1), derived in Appendix A),

$$= \frac{\text{Cov}(\text{CH}_2\text{O}_{\text{anthroVOCS}},\text{O}_3)}{\text{Var}(\text{CH}_2\text{O}_{\text{total}})} + \frac{\text{Cov}(\text{CH}_2\text{O}_{\text{BC}},\text{O}_3)}{\text{Var}(\text{CH}_2\text{O}_{\text{total}})} + \frac{\text{Cov}(\text{CH}_2\text{O}_{\text{biolSOP}},\text{O}_3)}{\text{Var}(\text{CH}_2\text{O}_{\text{total}})},$$
(1)

where Cov and Var denote covariance and variance, respectively. Equation (1) shows that the contribution of each CH_2O tracer, that is, the subslope values, to the O_3 - CH_2O regression slope is proportional to its covariance with O_3 . It is therefore possible to have both positive and negative slope contributions.

For the DISCOVER-AQ region during the Baltimore-Washington study, where the majority of CH₂O is biogenic, we can use the regression slope of O₃-CH₂O in the evaluation of model results using the observations. When extending the analysis using equation (1) to remote regions, the small variance of background CH₂O leads to abnormally large slopes, making it difficult to show the spatial distribution of the regression slope. We therefore use an *inversed slope* of Δ CH₂O/ Δ O₃ (equation (2), see the Appendix A) to illustrate the spatial distribution over the United States since the variance of O₃ is a more stable denominator that that of CH₂O,

 $\label{eq:least squares regression slope of CH_2O-O_3 } = \frac{Cov(CH_2O_{anthroVOCS},O_3)}{Var(O_3)} + \frac{Cov(CH_2O_{BC},O_3)}{Var(O_3)} + \frac{Cov(CH_2O_{biolSOP},O_3)}{Var(O_3)}.$

2.2. Surface Ozone Distribution Estimation

The total O₃ concentration is contributed by three major sources: anthropogenic O₃ production (O_{3anthro}), biogenic O₃ production (O_{3biolSOP}), and the transport from the lateral and upper model boundaries (O_{3background}; equation (3)). In the ozone estimation model, we estimate O_{3anthro} and O_{3biolSOP} using the regressions of O₃ with source-tagged CO and CH₂O. We will show in the next section that the correlations and regression slopes of O₃-CH₂O and CO-CH₂O are almost entirely due to biogenic isoprene over the eastern United States, where CH₂O concentrations are mostly due to oxidation of biogenic isoprene. We make use of this finding and use CH₂O as a proxy for O₃ related to biogenic emissions. We decompose surface O₃ concentrations in equation (3) into three components related to regional anthropogenic emissions, biogenic emissions, and background (not related to the emissions within the estimation domain). We approximate the emission related components using CO (equation (4)). Recognizing that we would like to make use of observation-based CO and CH₂O concentrations, we replace CO from anthropogenic emissions (CO_{anthro}) with CO-CO_{biogenic}-CO_{background}, where CO_{biogenic} as a function of CH₂O from biogenic isoprene oxidation (CH₂O_{biolSOP}; equation (5)). Considering that most of CH₂O is biogenic over the region in the summer, we replace CH₂O_{biolSOP} as CH₂O-CH₂O_{background} (equation (6)).

$$[O_3] = [O_3]_{anthro} + [O_3]_{bigenic} + [O_3]_{background}$$
(3)

$$\approx \frac{\Delta[O_3]}{\Delta[CO]_{anthro}} \times [CO]_{anthro} + \frac{\Delta[O_3]}{\Delta[CO]_{biolSOP}} \times [CO]_{biolSOP} + [O_3]_{background}$$
(4)

$$= \frac{\Delta[O_{3}]}{\Delta[CO]_{anthro}} \times \left([CO]_{total} - \frac{\Delta[CO]_{biolSOP}}{\Delta[CH_{2}O]_{biolSOP}} \times [CH_{2}O]_{biolSOP} - [CO]_{background} \right) \\ + \frac{\Delta[O_{3}]}{\Delta[CO]_{biolSOP}} \times \frac{\Delta[CO]_{biolSOP}}{\Delta[CH_{2}O]_{biolSOP}} \times [CH_{2}O]_{biolSOP} + [O_{3}]_{background}$$
(5)



$$\approx \frac{\Delta[O_3]}{\Delta[CO]_{anthro}} \times \left([CO]_{total} - \frac{\Delta[CO]_{biolSOP}}{\Delta[CH_2O]_{biolSOP}} \times \left([CH_2O]_{total} - [CH_2O]_{background} \right) - [CO]_{background} \right) \\ + \frac{\Delta[O_3]}{\Delta[CH_2O]_{biolSOP}} \times \left([CH_2O]_{total} - [CH_2O]_{background} \right) + [O_3]_{background}$$
(6)
$$= A \times [CO]_{total} + B \times [CH_2O]_{total} + C$$
(7)

where
$$A = \frac{\Delta[O_3]}{\Delta[CO]_{anthro}}$$
, $B = -\frac{\Delta[O_3]}{\Delta[CO]_{anthro}} \times \frac{\Delta[CO]_{biolSOP}}{\Delta[CH_2O]_{biolSOP}} + \frac{\Delta[O_3]}{\Delta[CH_2O]_{biolSOP}}$, $C = -\frac{\Delta[O_3]}{\Delta[CO]_{anthro}} \times [CO]_{background} + \left(\frac{\Delta[O_3]}{\Delta[CO]_{anthro}} \times \frac{\Delta[O_3]}{\Delta[CH_2O]_{biolSOP}} - \frac{\Delta[O_3]}{\Delta[CH_2O]_{biolSOP}}\right) \times [CH_2O]_{background} + [O_3]_{background}$.

We will show that this function works well for the eastern United States in the summer. The study domain (to be shown in Figure 4) is selected where >90% of monthly mean surface CH₂O is biogenic. Over the domain of surface O_3 estimation, we compute the $\frac{\Delta[O_3]}{\Delta[CO]_{anthro}}$, $\frac{\Delta[O_3]}{\Delta[CO]_{biolSOP}}$, and $\frac{\Delta[CO]_{biolSOP}}{\Delta[CH_2O]_{biolSOP}}$ values using the least squares regression slope formulations of O_3 -CO by Cheng et al. (2017) and O_3 -CH₂O in equation (1). The tagged-tracer simulations show that background values do not have a significant spatial dependence and we use a value of 60 ppbv for CO_{background} and a value of 200 pptv for CH₂O_{background}. Regression of equation (6) yields a value of 10 ppbv for O3background. As we will discuss in section 4, the unique feature of this surface O_3 estimation is that it is based on the temporospatial stability of the predicting parameters of the regression slopes and background values. To demonstrate the feasibility of this O₃ estimator, for hourly estimation from 11:00 a.m. to 4:00 p.m., we group the corresponding model data in all grid cells in the same hour and compute the regional regression slope values. Therefore, the estimated O_3 spatial variation is due to those of CO and CH₂O only. In order to test the reliability, the above estimation method is validated through leave-one-out cross validation (LOOCV). The method uses a single sample from the original data sets as the validation data, and the remaining samples (excluding the selected validation data point) are used in the estimator. Each sample in the data sets is used once as validation data. We conduct the validation individually for each hour of a day from 11:00 a.m. to 4:00 p.m. Of the 31 days from 1 to 31 July $(d_1, d_{2...}, d_{31})$, the day i (d_i) is selected as the validation data, and the remaining days (d_1 , d_2 ... $d_{(i-1)}$, $d_{(i+1)}$... d_{31}) are used to compute the regression slopes and estimate the surface O₃ distribution of day *i*. We exclude the data before 11:00 a.m. or after 4 p.m. when the estimation has large biases because photochemical production of $O_{3,i}$ CO, and CH₂O is still slow and the correlations among the species are not photochemically driven.

3. Results

3.1. Correlations and Regression Slopes of O₃-CH₂O and CO-CH₂O

We compare simulated O₃-CH₂O correlations to the DISCOVER-AQ observations as a function of altitude or as a function of time in Figure 1. Simulated and observed correlation coefficient (R) values are in good agreement. Strong correlation between O_3 and CH_2O is found from 300 m to the top of the boundary layer (~2.5 km) with a narrow range of R values (~0.75). Near the surface, simulated R values show a slight decrease due to the increase of the contribution of surface primary emissions of CH₂O, which are not directly related to photochemical O_3 production. From the upper boundary layer to the free troposphere, R values show a drastic drop and changes sign from positive to negative. The sign change reflects the rapid decrease of CH₂O with altitude in the lower free troposphere where O_3 increases with altitude (Figure S2 in the supporting information). In the middle and upper boundary layers, the relative contribution by photochemical CH₂O production increases with altitude. Therefore, the concurrent photochemical production of O₃ and CH₂O is a major factor contributing to the observed positive correlation between O₃ and CH₂O in the boundary layer. We also compare the simulated and observed O₃-CH₂O correlation coefficient as a function of time of the day with data from 300 m to 2.5 km (Figure 1b). These R values represent the spatial correlation in a given hour and are somewhat lower than the spatial correlations in the vertical just discussed. The model is in good agreement with the observations except the underestimation at 9 a.m. and overestimation at 5 p.m., which again is related to the fact that photochemistry becomes less important here.

The observed regression slope of O_3 to CH_2O is also captured by model simulation with satisfactory agreement in both vertical distribution and diurnal variation (Figures 1c and 1d). The O_3 -CH₂O regression slope at a given altitude is underestimated by the model at 0.3–2.5 km by ~15%. However, the observed slight





Figure 1. Observed and simulated O_3 -CH₂O correlation coefficients (*R*) and regression slope and subslopes (equation (1)) as a function of altitude (of data for 11 a.m. to 4 p.m.; a, c) and as a function of local time (of data for altitude of 0.3–2.5 km; b, d). The subslopes due to varied CH₂O sources are shown using areas filled with different colors (equation (1)). The legends for different CH₂O sources are the same as in equation (1). The horizontal bars in (c) and vertical bars in (d) show the observed or simulated standard deviations of the regression slopes. The *R*, slope, and subslope values are computed using the DISCOVER-AQ (DAQ) observations or corresponding model data at a given altitude bin or for a given period. DISCOVER-AQ = Deriving Information on Surface Conditions from Column and Vertically Resolved Observations Relevant to Air Quality; REAM = Regional chEmical trAnsport Model.

increase (from ~5 to ~8 ppbv/ppbv) in the boundary is captured by the model. The observed O_3 -CH₂O regression slope at a given hour is better simulated than that at a given altitude. We further quantify source contributions by decomposing the O_3 -CH₂O regression slope into the different CH₂O sources using equation (1). During DISCOVER-AQ, the regression slope of O_3 -CH₂O is contributed almost exclusively by that of biogenic CH₂O.

Over the United States, we use the model results in Figure 2 to understand the relative contributions of CH_2O sources to the CH_2O-O_3 regression slope near the surface (equation (2)). The slope decomposition results show that the contribution from biogenic isoprene to the slope of O_3 to CH_2O is overwhelming over most regions of the eastern United States. Over the regions where it dominates, biogenic CH_2O has positive correlations with O_3 due to the concurrent production of these two species from the oxidation of biogenic VOCs. The exception is in central United States, where biogenic CH_2O concentrations are high but O_3 concentrations are low due to low NO_x concentrations. The chemical loss of O_3 leads to a negative regression slope, which appears high because of the low variance of O_3 in the region (equation (2)). Without significant photochemical production, the variation of O_3 in this region is low and is therefore relatively easier to estimate than the other regions. Overall, the short lifetimes of biogenic isoprene and CH_2O make it possible to use its concentrations to estimate the spatial variation of O_3 using relationships like equation (7).

We also compare observed and simulated correlation and regression slopes of CO and CH₂O using an equation similar to equation (1) for the Baltimore-Washington area. This information is also used in equation (7). The simulation results are in good agreement with the observations as a function of altitude or time of a day (Figure 3). Strong correlation between CO and CH₂O (R = ~0.85) is observed and simulated from 300 m to 2.5 km (Figure 3a). This correlation is most likely due to the coemissions from anthropogenic sources and the coproduction of mostly biogenic CO and CH₂O. In this altitude range, the regression slope of CO to CH₂O is about ~20 ppbv/ppbv without little variation between 9 a.m. and 5 p.m. (Figure 3c). Below 300 m, simulated *R* value decreases toward the surface because the contribution of CO primary emissions increases significantly near the surface while the secondary formation is still the major source of CH₂O (Figure S3 in the supporting information). These two processes are not correlated. Correspondingly, the slope of CO to CH₂O increases from 200 m to surface because the gradient of CO is larger than that of CH₂O (Figure S2) due to the

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Figure 2. Distribution of the regression slopes and subslopes of surface CH_2O-O_3 (equation (2)) over the United States. The title *All* stands for the slope computed by total O_3 and total CH_2O concentration; the titles of *Bio ISOP, Anthro VOCs*, and *BC* stand for the subslopes computed by total O_3 and CH_2O from primary anthropogenic emissions and oxidation of anthropogenic VOCs, oxidation of biogenic isoprene, and transport from the lateral and upper model boundaries, respectively. The slope and subslope values on each grid are computed using selected hourly data on daytime of 11: 00 a.m. to 04: 00 p.m. from 1 to 30 July. VOCs = volatile organic compounds.



Figure 3. Observed and simulated CO-CH₂O correlation coefficients (*R*) and regression slope for the Baltimore-Washington area as a function of altitude for daytime of 11:00 a.m. to 04:00 p.m. (a, c) and local time for altitude of 0.3-2.5 km (b, d). The horizontal bars in (c) and vertical bars in (d) show the observed standard deviations of the regression slopes. Shaded blue areas in (c) and (d) show simulated standard deviations of the regression slopes. The *R* and slope values are computed using the DISCOVER-AQ (DAQ) observations or corresponding model data at a given altitude bin or for a given time period. DISCOVER-AQ = Deriving Information on Surface Conditions from Column and Vertically Resolved Observations Relevant to Air Quality; REAM = Regional chEmical trAnsport Model.

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Figure 4. Distributions of monthly mean (11:00 a.m. to 4:00 p.m.) REAM and LOOCV estimation of surface O_3 concentrations for July 2011 under different anthropogenic emission scenarios. In LOOVC hourly estimation, the regional parameters in equation (7) are estimated using data not including the day of estimation. Scatterplots of corresponding grid-cell hourly REAM and LOOCV estimation data are shown in the third row; the 1:1 line is shown in red. Seven emission scenarios are presented. +50%, +30%, +15%, Standard, -15%, -30%, and -50% on the top of columns denote 150%, 130%, 115%, original, 85%, 70%, and 50% of anthropogenic CO and NO_x emissions of the 2011 National Emission Inventory. The comparison statistics are listed in Table 1. REAM = Regional chEmical trAnsport Model; LOOCV = leave-one-out cross validation

surface primary emissions. However, we also test the correlation of surface $CO_{biogenic}$ and CH_2O , which shows an *R* value in a range of 0.85 to 0.95 in the areas where more than 90% monthly mean CH_2O is from biogenic isoprene oxidation. It implies that CH_2O and the correlation of $CO-CH_2O$ can be used to separate anthropogenic and biogenic CO. Above 2.5 km, the *R* value decreases from the boundary layer top to the free troposphere. This is because above the boundary layer CO is mainly from transport from lateral and upper boundaries, which does not contribute to CH_2O as much due to its short chemical lifetime (Figure S3). The slope of CO to CH_2O also increases from the boundary layer to the free troposphere due to low concentrations of CH_2O in free troposphere. As a function of time, the *R* value does not show a significant variation. The regression slope of CO-CH_2O remains at ~20 ppbv/ppbv in daytime except higher values in the morning (before 8: 00 a.m.) when photochemistry is weak.

3.2. Surface Ozone Estimation Using Equation (7)

For July 2011, we estimate the LOOCV surface O_3 distribution using equation (7). The averaged parameters of equation (7) for the 30 validations are listed in Table S2 in the supporting information. To examine the sensitivity of the estimation to emissions, we also increase (or decrease) anthropogenic emissions by 15%, 30%, and 50%, respectively. The full REAM model is run with different emissions. For the LOOCV estimation

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Table 1 Comparison of Hourly REAM and LOOCV Estimation of Surface O₃ Concentrations Under Different Anthropogenic **Emissions Scenarios RFAM** mean Mean **Emission** Estimation mean Least squares R^2 ±Std (ppbv) MSE slope (ppbv/ppbv) scenarios ±Std (ppbv) Bias +50% 7.6 0.90 63.0 ± 8.2 64.1 ± 8.1 1.1 1.06 0.9 0.89 1.04 +30% 59.0 ± 7.8 59.9 ± 7.6 7.1 +15% 53.9 ± 7.5 54.6 ±7.5 0.7 8.5 0.87 1.02 Standard 51.4 ±7.0 51.8 ± 6.9 0.4 7.8 0.86 0.99 9.5 0.97 -15% 48.2 ± 6.8 47.2 ±6.8 -1.00.81 -30% 43.2 ± 6.3 11.0 0.77 0.95 44.3 ± 6.6 -1.1-50% 38.8 ± 6.2 37.3 ± 5.5 -1.5 13.6 0.65 0.91

Note. All grids in the domain in all certain hours (11:00 a.m. to 4:00 p.m.) are grouped together. REAM = Regional chEmical trAnsport Model; LOOCV = leave-one-out cross validation; Std = standard deviation; MSE = mean squared error.

using equation (7), the $\frac{\Delta[\mathbf{O}_3]}{\Delta[\mathbf{CO}]_{anthro}}$, $\frac{\Delta[\mathbf{O}_3]}{\Delta[\mathbf{CO}]_{biolSOP}}$, and $\frac{\Delta[\mathbf{CO}]_{biolSOP}}{\Delta[\mathbf{CH}_2\mathbf{O}]_{biolSOP}}$ values are the same as in the standard simulation. The only changes are for the CO and CH₂O surface concentrations used. The premise is that the regression slopes of $\frac{\Delta[\mathbf{O}_3]}{\Delta[\mathbf{CO}]_{ahthro}}$, $\frac{\Delta[\mathbf{O}_3]}{\Delta[\mathbf{CO}]_{biolSOP}}$, and $\frac{\Delta[\mathbf{CO}]_{biolSOP}}{\Delta[\mathbf{CH}_2\mathbf{O}]_{biolSOP}}$ are relatively stable with respect to emission changes and therefore the estimation model using equation (7) provides a robust means to estimate surface O₃ distribution when the distributions of CO and CH₂O are known.

Since equations (4)–(7) are more accurate when CH₂O is dominated by oxidation of biogenic isoprene, the estimation evaluation is only for the eastern U.S. regions where monthly mean biogenic CH_2O is >90%. We compare the averaged estimation results with REAM results for these regions under different emission scenarios in Figure 4. In the standard simulation, the estimation shows a similar distribution and explains 86% of the variance of the full REAM results ($R^2 = 0.86$) with no significant overall bias (Table 1), although the scatterplot shows a slight tendency of low biases for O_3 above 60 ppbv with an average of -1.7 ppbv. As the anthropogenic emissions decrease by 15–50%, the REAM model shows a decrease from 3.0 to 12.4 ppbv on average in these regions. The estimation model using equation (7) overestimates the O_3 decrease by 1.0–1.5 ppbv on average and the explained variance decreases from 81% to 65%. When anthropogenic emissions increase by 15%–50%, the full REAM results show surface O₃ increases from 2.7 to 11.8 ppbv. As anthropogenic emissions increase, the O₃ hot spots due to urban emissions become more obvious. The estimation model using equation (7) shows similar features but overestimates the O₃ increase by 0.7–1.1 ppbv on average. However, the explained variance (R^2 value) increases from 0.87 to 0.90 due largely to good estimations of urban increases. We also tested the estimation model performance by increasing (or decreasing) biogenic isoprene emissions by 15%, 30%, and 50%, respectively. The full REAM model shows much lower dependence of surface O_3 to biogenic emissions than anthropogenic emissions (Figure S4 in the supporting information). The estimation model shows similar results with R^2 values ranging from 0.77 to 0.86 (Table S1 in the supporting information).

4. Discussion

The surface O₃ estimator (equation (7)) works very well for the regions shown in Figure 4, and, it is also quite robust with R^2 values >0.77 for anthropogenic emissions in the range of -30% to +50% and biogenic emissions of -50% to +50%. Therefore, it has the potential of being used for rapid O₃ distribution assessment if the distributions of surface CO and CH₂O are known. Given the current lack of usable surface CO and CH₂O observations, we cannot test the estimator using observed data. In this analysis, the key parameters of the estimator, that is, the regression slopes of $\frac{\Delta[O_3]}{\Delta[CO]_{anthro}}$, $\frac{\Delta[O_3]}{\Delta[CO]_{biolSOP}}$, and $\frac{\Delta[CO]_{biolSOP}}{\Delta[CH_2O]_{biolSOP}}$ are based on model simulations, although we show that model simulations are in good agreement with DISCOVER-AQ observations (Figures 1 and 3; Cheng et al., 2017). If the distributions of O₃, CO, and CH₂O are known, equation (6) can be used to obtain the observation-based regression slopes using least squares regression. These parameters can provide insights in understanding of biases of air quality model simulations and be applied to improve the model.

Surface CO measurements are readily obtained if existing instruments are calibrated and the reporting limit is lowered. In situ observation of CH₂O is more complex and expensive than CO. For both species, the high concentrations in the boundary layer and very low concentrations in the stratosphere imply that satellite instruments have better sensitivity to derive their near-surface concentrations than O₃. The relatively large uncertainties of the current generation instruments on Sun-synchronous satellites can be greatly reduced (due in part to the large increase of observation frequency) by instruments on board geostationary satellites such as TEMPO over North America (Chance et al., 2013), SENTINEL-4 over Europe (Ingmann et al., 2012), and GEMS over East Asia (Bak et al., 2013). High-quality CH₂O measurements over the United States will be available from TEMPO. Deployment of near-IR and thermal instruments on geostationary satellites will be needed for improved satellite measurements of near-surface CO. More complex statistical methods can be applied to combine such derived O₃ estimation with in situ surface O₃ observations (Y. Zhang et al., 2018).

The estimator is valid only in the regions where CH_2O is dominated by oxidation of biogenic isoprene during the period of the year when isoprene emissions are large. As anthropogenic emissions are expected to continue decreasing (Cheng et al., 2017), the regions where biogenic CH_2O dominates and the estimator can be applied will increase. There are limitations in the estimator of equation (7). The relatively good performance of the estimation model with changing anthropogenic emissions is not because surface O_3 is insensitive to anthropogenic NO_x emissions. In fact, most of the changes shown in Table 1 are due to NO_x . However, surface NO_x is not as useful a predictor as CO and CH_2O in the formulation of equation (7). O_3 is a secondary pollutant while, NO_x is mostly a primary pollutant in our study region. In comparison, biogenic CO and CH_2O are secondary. As NO_x is oxidized and its concentration decreases, O_3 , biogenic CO, and CH_2O are produced and their concentrations increase. The anthropogenic CO is a better tracer for the cumulative effect of O_3 production by anthropogenic NO_x because of its much longer lifetime than NO_x and fast-reacting VOCs (e.g., Chin et al., 1994).

In REAM simulations, we make an implicit assumption that NO_x and CO emission ratios of anthropogenic sources do not change. This assumption is not always valid; for example, NO_x emission reduction from the electric generating utility sources in the past two decades did not reduce CO emissions significantly (https://www.epa.gov/air-emissions-inventories/air-pollutant-emissions-trends-data). When the emission changes of anthropogenic NO_x and CO are sufficiently different, the estimator needs to be reconstructed with updated regression slopes of $\frac{\Delta[O_3]}{\Delta[CO]_{biotSOP}}$, and $\frac{\Delta[CO]_{biotSOP}}{\Delta[CO]_{biotSOP}}$. Furthermore, if a region is affected by pollutant sources such as fires that are not included in the model, the estimation results will be biased. Another uncertainty of this estimator could come from isoprene chemistry. If a specific isoprene oxidation pathway significantly affects the regression slopes of $\frac{\Delta[O_3]}{\Delta[CO]_{biotSOP}}$ and $\frac{\Delta[O_3]}{\Delta[CO]_{biotSOP}}$, and $\frac{\Delta[CO]_{biotSOP}}{\Delta[CO]_{biotSOP}}$, equation (6) can be applied to derive these parameters on the basis of the observations, which can then be applied to evaluate the chemical pathway representation in the model.

5. Conclusions

Extensive measurement of O_3 , CO, and CH_2O were conducted during the DISCOVER-AQ aircraft experiment in July 2011 over the Washington-Baltimore area. We find strong correlation and stable linear regression slopes of O_3 -CH₂O, O_3 -CO, and CO-CH₂O with no significant variation with time (11 a.m. to 4 p.m.) or altitude in the boundary layer. The concentrations, correlations, and regression slopes of these tracers are reproduced well by the REAM model. We find that biogenic isoprene oxidation makes most of the contribution to the regression slopes of CH₂O-O₃ in large regions of the eastern United States using the slope decomposition method by tracing separately three different CH₂O sources, including primary anthropogenic emissions and oxidation of anthropogenic VOCs, oxidation of biogenic isoprene, and transport from the lateral and upper model boundaries.

Making use of the robust regression slopes, we construct a surface ozone estimation model using the distributions of CH_2O and CO as input parameters. In this model, CH_2O is used as a proxy to calculate O_3 and CO produced by the oxidation of biogenic VOCs. The estimator can explain >77% of the surface O_3 variance simulated by the full 3-D model in the range of 70% to 150% of the anthropogenic emissions. It provides a fast regional surface O_3 estimation in most regions of the eastern United States in summer where CH_2O is



dominated by oxidation of biogenic isoprene. With high-quality geostationary satellite observations of CO and CH₂O, the estimator could be applied to improve surface O₃ distribution, which is challenging to measure directly from space. The function of equation (6) relates surface concentrations of O_3 to those of CO and CH₂O. Using a sizable observation data set of these concentrations of a given region, regional metrics, such as the regression slopes of $\frac{\Delta[O_3]}{\Delta[CO]_{anthor}}$, $\frac{\Delta[O_3]}{\Delta[CO]_{biolSOP}}$, and $\frac{\Delta[CO]_{biolSOP}}{\Delta[CH_2O]_{biolSOP}}$, can be empirically determined and applied to investigate model performance and biases. Therefore, the estimation model provides the means of using observations to evaluate and improve full-fledged 3-D air quality models.

Appendix A

CH₂O concentration is the sum of that from primary anthropogenic emissions and the oxidation of anthropogenic VOCs (CH₂O_{anthroVOCs}), the oxidation of biogenic isoprene (CH₂O_{biolSOP}), and transport from model lateral and upper boundaries (CH₂O_{BC}):

$$[CH_2O]_{total} = [CH_2O]_{anthroVOCs} + [CH_2O]_{BC} + [CH_2O]_{biolSOP}.$$
(A1)

The slope of O₃ to CH₂O in a least squares regression is thus

$$\frac{\left(\left[\mathsf{CH}_{2}\mathsf{O}\right]_{\mathsf{total}}-\overline{\left[\mathsf{CH}_{2}\mathsf{O}\right]_{\mathsf{total}}}\right)\left(\left[\mathsf{O}_{3}\right]-\overline{\left[\mathsf{O}_{3}\right]}\right)}{\mathsf{Var}(\mathsf{CH}_{2}\mathsf{O}_{\mathsf{total}})} \tag{A3}$$

$$= \frac{\left(\left[\mathsf{CH}_2\mathsf{O}\right]_{\mathsf{anthroVOCs}} + \left[\mathsf{CH}_2\mathsf{O}\right]_{\mathsf{BC}} + \left[\mathsf{CH}_2\mathsf{O}\right]_{\mathsf{biolSOP}} - \overline{\left(\left[\mathsf{CH}_2\mathsf{O}\right]_{\mathsf{anthroVOCs}} + \left[\mathsf{CH}_2\mathsf{O}\right]_{\mathsf{BC}} + \left[\mathsf{CH}_2\mathsf{O}\right]_{\mathsf{biolSOP}}\right)}\right) \left(\left[\mathsf{O}_3\right] - \overline{\left[\mathsf{O}_3\right]}\right)}{\mathsf{Var}(\mathsf{CH}_2\mathsf{O}_{\mathsf{total}})}$$

$$= \frac{\overline{\left(\left[\mathsf{CH}_{2}\mathsf{O}\right]_{anthroVOCs} - \overline{\left[\mathsf{CH}_{2}\mathsf{O}\right]_{anthroVOCs}}\right)\left(\left[\mathsf{O}_{3}\right] - \overline{\left[\mathsf{O}_{3}\right]}\right)}{+\overline{\left(\left[\mathsf{CH}_{2}\mathsf{O}\right]_{biolSOP} - \overline{\left[\mathsf{CH}_{2}\mathsf{O}\right]_{biolSOP}}\right)\left(\left[\mathsf{O}_{3}\right] - \overline{\left[\mathsf{O}_{3}\right]}\right)}}{Var(\mathsf{CH}_{2}\mathsf{O}_{total})}$$
(A5)

$$= \frac{\text{Cov}(\text{CH}_2\text{O}_{anthroVOCs}, \text{O}_3)}{\text{Var}(\text{CH}_2\text{O}_{total})} + \frac{\text{Cov}(\text{CH}_2\text{O}_{BC}, \text{O}_3)}{\text{Var}(\text{CH}_2\text{O}_{total})} + \frac{\text{Cov}(\text{CH}_2\text{O}_{biolSOP}, \text{O}_3)}{\text{Var}(\text{CH}_2\text{O}_{total})},$$
(A6)

where \overline{X} donates the average value of X.

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